

# Modeling Mississippi River dredging strategies after the lock closure at Upper St. Anthony Falls

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## Abstract

With the closure of the Upper St. Anthony Falls (USAF) lock in 2015, new opportunities have arisen to investigate eliminating channel maintenance in the upper navigation pools of the Mississippi River near Minneapolis and St. Paul, Minnesota. Since lockages through USAF have halted, the current dredging activities that are in place to ensure a nine-foot navigation draft in the pool may no longer be required. Additionally, due to reduced commercial boat traffic traveling to USAF through Lock & Dam No. 1 (LD1), elimination of channel dredging through this reach may also be warranted. The results of this sediment transport modeling study, using HEC-RAS hydraulic modeling software, show the relative differences in dredging quantities between the current channel maintenance practices and proposed alternatives. Modeled impacts for two alternatives, eliminating dredging above USAF and eliminating dredging above LD1, are quantified for the Mississippi River system through Lake Pepin in order to assess the viability of each strategy. In addition, future studies may utilize this model to analyze sediment trends through Lake Pepin and to investigate the feasibility of major operational changes (e.g. water level drawdowns) or physical changes (e.g. dam modification) at the structures in the navigation system.

## Introduction

### Background

The Water Resources Reform and Development Act of 2014 (WRRDA 2014) required that the Upper St. Anthony Falls Lock and Dam (USAF Lock) be permanently closed. This closure occurred on June 10th, 2015. The closure of the USAF Lock essentially ended the need for annual maintenance dredging in the commercial navigation channel in Pool 1 and the Upper St. Anthony Falls Pool. The decision that has to be made by the St. Paul District of the U. S. Army Corps of Engineers (USACE) is whether to 1) continue channel maintenance as usual, 2) stop channel maintenance in the USAF Pool and in Pool 1, or 3) develop a sediment management strategy based on the beneficial use of dredge material. Because navigation channel dredging can be a large sink for sand-size sediment, reducing dredging in the USAF Pool and Pool 1 may eventually have an effect on downstream reaches.

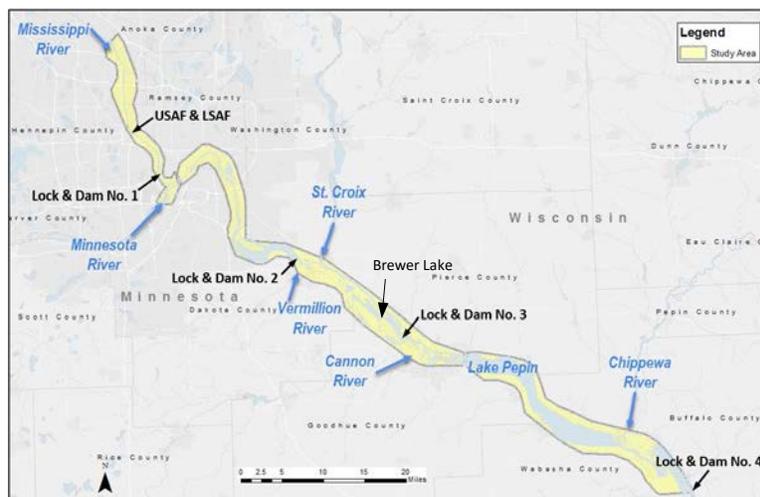
The St. Paul District of the U.S. Army Corps of Engineers (USACE) is responsible for maintaining a 9-foot navigation channel on the Upper Mississippi River (UMR) between Minneapolis, Minnesota and Guttenburg, Iowa. This includes the lower 14.7 miles of the Minnesota River, and portions of the lower St. Croix and Black Rivers. The Upper St. Anthony Falls Lock is the upstream most navigation dam on the Mississippi River and the head of navigation is just a few miles upstream of the lock. Maintaining the 9-foot channel is done through periodic dredging and through a system of locks and dams. The USACE dredges and disposes of approximately 66,000 cubic yards of sand annually between the Upper St. Anthony Falls Pool (USAF) and Lock and Dam 1, and 160,000 cubic yards annually on the UMR between Lock and Dam 1 and Lake Pepin. The total from both reaches represents over 25-percent of the district-wide dredging. In addition to the cost associated with channel maintenance dredging, other sediment related impacts in this reach include a turbidity impairment (MPCA, 2012), off-

channel sediment deposition affecting habitat and recreational boating, reduced light penetration and aquatic vegetation growth, and accelerated sediment deposition in Lake Pepin. It is estimated that 85 - 90 percent of the sediment deposited in Lake Pepin originates from the Minnesota River watershed (Engstrom et. al., 2009). Figure 1 shows the extent of the study area.

To estimate the effects of navigation channel dredging, off-channel sediment deposition, and tributary sediment loads on sediment transport on the UMR, the USACE developed a district-wide bed material sediment budget in 2003 (USACE, 2003). Bed material refers to sand-size sediment that can be found on the bed of the main channel, but can be transported as bed load or suspended load. This bed material budget was based on interpretation of available sediment transport information at U.S. Geological Survey (USGS) gaging stations, long-term channel dredging data, studies of sediment transport and deposition, and measured hydraulic characteristics on the UMR. Total sediment load measurements obtained on the Minnesota River at Ft. Snelling during the years 2011 to 2015 (Groten et. al., 2016) have improved the sediment budget significantly. However, while the sediment budget has been a valuable tool, it isn't a numerical model and can't predict the temporal and spatial effects of changed sediment transport capacity and sediment loads.

## Project Location and Study Area

The study area is on the Mississippi River 9-Foot Navigation Channel between River Mile (RM) 857.6, the upstream limit of the 9 foot channel project, and RM 764, the downstream end of Lake Pepin. For hydraulic modeling purposes, the upstream extent has been extended to RM 866 to include the Anoka gage on the Mississippi River and the downstream extent has been extended to RM 753 to capture the downstream control of the water level for Lake Pepin at Lock and Dam No. 4. This reach includes numerous structures and incoming tributaries, shown in Figure 1.



**Figure 1.** Overview of the Modeling Study Area

## Purpose and Need

A numerical model is needed for the reach of the Mississippi River from the USAF Pool to Lake Pepin to simulate the effects of changed dredging in the USAF Pool and Pool 1. The primary purpose of the model is to simulate the spatial and temporal effects of dredging changes in USAF and Pool 1 on downstream dredging and backwater deposition of sand sized sediment in

pools 2, 3, and 4. Other purposes include determining the effects of a secondary channel closure proposed for the Brewer Lake Inlet in Pool 3. The appropriate model must be capable of modeling the complexities of flow exchanges between main channel and backwater areas and advanced operations of multiple lock and dam structures, while also being capable of modeling long reaches of river over 100 miles in length. Advanced two-dimensional and three-dimensional models would be appropriate for capturing complex hydraulic behavior, but would not be efficient over a domain as large as the proposed study area. Conversely, simple spreadsheet models and sediment budgets would be efficient, but incapable of capturing the hydraulic complexities of this system. The HEC-RAS one-dimensional model has been selected for this study as appropriate as it is capable of effectively modeling hydraulics and sediment over large domains as well as capturing smaller scale complexities at flow splits and structures.

## Methods

### Data Collection

**Flow and Stage Gage data:** Water surface elevation data, flow records, and sediment measurements are important pieces of data for both the construction and calibration of a hydraulic and sediment model. Water surface elevation data is available through continuous measurements using Data Collection Platform (DCP) instruments and through daily observations of stage data at structures, points of interest, and established gage locations. The U.S. Army Corps of Engineers collects continuous and daily records of water surface elevation for pool and tailwater (TW) levels at each of the operated lock and dam structures as well as at “control point” locations which are used for hinge-point operations of the navigation system. The United States Geological Survey (USGS) collects water surface elevation at established gaging stations which can be converted to a continuous record of discharge or streamflow by maintaining a stage-discharge relationship for each gage location through the periodic measuring of discharge at that location (Olson & Norris, 2007). The most complete shared record for all the gages in the study area is the period from 2007-2015.

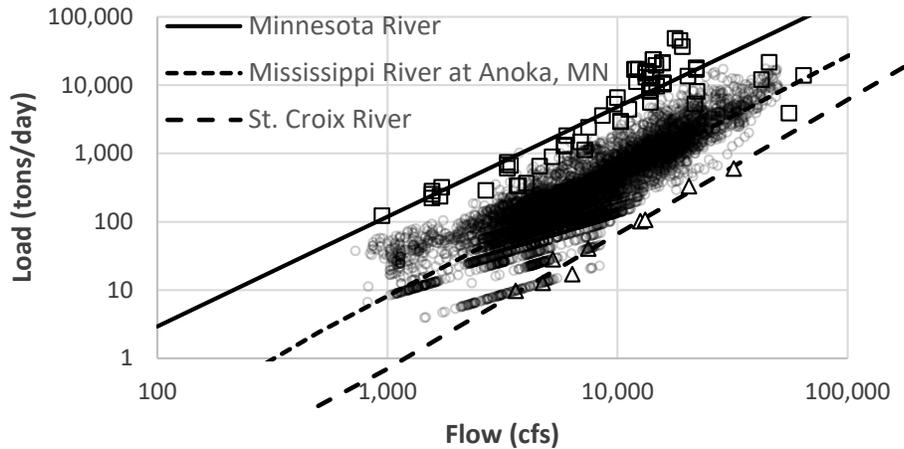
**Suspended Sediment:** In addition to measurements of stage and flow at various gage locations, the USGS collects field samples of suspended sediment concentration and sediment grain size distribution for use in water quality and runoff analyses. This suspended sediment data can also be used as an input to a sediment transport model. The units for the collected concentration values are recorded in mass per volume, or typically milligrams per liter using the International System of Units (SI). The HEC-RAS sediment model requires total sediment load in units of weight per time, or tons per day using English units. To convert the concentration to a total load, the concentration needs to be multiplied by the instantaneous river flow that occurs at the time of the concentration measurement, as well as a coefficient to convert to the appropriate units. The total load, in tons per day, can be calculated by the following equation (Porterfield, 1972):

$Q_s = Q_w * C_s * K$  where

$Q_s$	=	Sediment discharge or sediment load, in tons per day (tons/day)
$Q_w$	=	Discharge or streamflow, in cubic feet per second (ft <sup>3</sup> /s or cfs)
$C_s$	=	Concentration of suspended sediment, in milligrams per liter (mg/L)
$K$	=	0.00269, the coefficient to convert units

$$K = \left( \frac{86,400 \text{ seconds}}{1 \text{ day}} \right) * \left( \frac{1 \text{ meter}}{3.28 \text{ feet}} \right)^3 * \left( \frac{1000 \text{ liters}}{1 \text{ cu. meter}} \right) * \left( \frac{1 \text{ kilogram}}{10^6 \text{ milligrams}} \right) * \left( \frac{2.2 \text{ pounds}}{1 \text{ kilogram}} \right) * \left( \frac{1 \text{ ton}}{2000 \text{ pounds}} \right) = 0.00269$$

Suspended sediment concentration measurements are available for the three main inflows to the model domain: the Mississippi River, the Minnesota River, and the St. Croix River. The Mississippi River has a total of 7,714 sediment measurements while the Minnesota River has a total of 74 measurements and the St. Croix River has 9 observations that can be used to develop a flow-load curve. A power-fit regression of the log-transformed values of flow and load was developed for each set of data and used as an initial best estimate for the flow-load relationship. The measured data and the best estimate curves are shown in Figure 2.



**Figure 2.** Flow-Load Relationships for three major inflows to the sediment model

The flow-load relationships for these three major inflows show roughly two orders of magnitude difference in total sediment load. For example, at 10,000 cfs the best estimate for the St. Croix River is 77 tons/day, the best estimate for the Mississippi River is almost ten times greater at 486 tons/day, and the best estimate for the Minnesota River is ten-fold greater still at 6061 tons/day. The higher sediment loads that are found within the Minnesota River Basin can help explain why this tributary contributes over 85% of the sediment that makes its way to Lake Pepin (Engstrom, 2009).

Sediment samples collected by the USGS, in addition to obtaining a measurement of concentration, can be analyzed to determine the sediment grain size distribution. The percentage of the suspended sediment that falls into the various grain classes of sands (0.0625-1 mm), silts (0.004-0.0625 mm), and clays (< 0.004 mm) can be determined through sieve and hydrometer tests as described in the American Society for Testing and Materials (ASTM) procedure D422-63(2007)e2 (ASTM, 2007). There is an inherent amount of variability in the testing for particle size distribution which is difficult to capture in a 1D sediment model. For this reason, most of the inflows in the model were assumed to have the same suspended sediment gradation based off of the median values from the numerous samples. The one exception, the Minnesota River, was assumed to have a higher percentage of finer material (coarse silts and fine sands) based on the median values of samples from that collection site.

**River Bed Gradations:** The sediment model allows for different gradations of bed material to be assigned at each cross-section in the model. Bed samples were found throughout the model domain area, both on the main channel and in backwater areas such as marinas. The various types of bed gradations were sorted in groups based on pool and flow type (i.e. main channel vs. backwater areas or sloughs). The median values were taken for each group and applied to the various reaches as appropriate. For example, the North & Sturgeon Lake area was modeled with bed gradations for “Pool 3 Coulee/Sloughs”.

Ultimately, since the system is primarily depositional rather than erosional, the sediment modeling results are not very sensitive to the bed gradations.

**Digital Elevation Model (DEM):** The modeling domain must extend far enough upstream to encompass the dredge locations above Upper & Lower St. Anthony Falls (USAF & LSAF) and far enough downstream to create a downstream boundary that does not affect the stage and flow calculations at Lake Pepin. The U.S. Army Corps of Engineers St. Paul District has numerous years of extensive bathymetric datasets in each of the pools in the study area through surveys performed by USACE for dredging, navigation, and ecosystem restoration purposes. The St. Paul District GIS Section has merged these datasets with above-low-water Light Detection and Ranging (LiDAR) data collected by the Minnesota Department of Natural Resources (MN DNR) since 2008; providing seamless datasets for pools throughout the study area. These datasets have been merged for this study to create a single Digital Elevation Model (DEM), which is used to attribute elevation data to the hydraulic model features.

## Model Construction

The selected software for the modeling effort is HEC-RAS (USACE, 2016). This software, originally developed as a one-dimensional (1D), steady-flow hydraulic modeling software package, now has capabilities for unsteady flow, sediment modeling, and two-dimensional (2D) flow. For this effort, the software is used to construct a 1D, unsteady-flow, hydraulic river model; calibrate the hydraulic model to collected stage and flow data; further develop the model into a 1D, unsteady-flow, *sediment* hydraulic river model; calibrate the sediment model to observed dredging records; and assess sediment impacts for changes to existing system operations. HEC-RAS does not currently have the capabilities to model sediment in 2D. Instead a 1D model, which calculates the water surface profile by solving the 1D Saint-Venant (momentum) equations over successive river “cross-section” features, is used rather than a 2D model capable of solving the 2D Saint-Venant equations or diffusive wave equations across multidirectional cell features. While a 1D model is less detailed in nature, it can provide shorter model run times for added complexities such as sediment, multi-year flow records, and large model domains.

The cross-section location data, river centerline features, Manning’s n-values for roughness, and ineffective flow limits were taken from various existing HEC-RAS models developed for the Mississippi River in this area:

- Mississippi River through St. Paul (Pool 2) developed as part of a USGS study (Czuba et. al. 2014)
- Lower Minnesota River from latest Corps Water Management System (CWMS) Modeling by USACE, St. Paul District in 2016
- Mississippi River through Pools 3 & 4, developed as part of a modeling effort for the Nuclear Regulatory Commission in 2015

While the 1D model cannot capture the complexities of two-dimensional flow, the floodway can be modeled as multiple channels to better capture the flow splits near Grey Cloud Island (Baldwin Lake and Spring Lake), Prairie Island (Vermillion River and North & Sturgeon Lakes), and Red Wing (Wisconsin Channel).

The cross-section and “lateral structure” features that connect the various reaches are “cut” from the developed seamless DEM to ensure that model represents the conditions with the best available data. The lock and dam structures are imported from the previously developed models to ensure that the gates, sills, and dam crests were set to the appropriate sizes, elevations, and datum.

All elevations used in the modeling effort and presented in this report are in North American Vertical Datum of 1988 (NAVD 88). The conversion from the National Geodetic Vertical Datum of 1929 (NGVD29) is to add 0.194 feet at the upstream end of study area and 0.036 feet at the downstream end of study area.

## **Hydraulic Model Calibration and Validation**

The hydraulic model was calibrated to water surface elevation data at pool, tailwater, and control point gages and to flow estimates at USGS gage locations. The metric used to assess the calibration to observed flow is the Nash-Sutcliffe model efficiency coefficient (NSE) which is a common metric used to assess the predictive power of hydrologic models (Nash & Sutcliffe, 1970). The model accuracy is high as the NSE values approach a value of 1. The four discharge gages that were compared showed NSE values of 0.95-0.99 indicating that the model is very accurate in terms of flow.

Backwater flow was also validated against periodic backwater measurements collected by the St. Paul District. Various locations throughout Pool 2 and Pool 3 were measured to help estimate the flow conveyance of the main channel compared to backwater or side channel areas. In Pool 2, lower velocity areas such as Baldwin Lake and Spring Lake still convey up to 20% of the total flow on the river. In Pool 3, the Vermillion River and North & Sturgeon Lakes can convey an even greater percentage of the total flow. At Pool 3 in particular, the flow splits to secondary channels and backwater lakes are very complex, with numerous sloughs and breakout areas allowing for interchanging flow. The 1D model is able to capture the flow splits surprisingly well, with strong validation between the modeled flow and the periodic measurements of flow.

The modeled water surface elevation data at the navigation structures and control points was compared to the observed data using the estimator of mean square error (MSE) which is the sum of the squared difference between observed and predicted values (Legates & McCabe, 1999). This is another common metric in statistical modeling for goodness of fit, with values closer to 0 indicating higher accuracy. Values at the various gages generally range from 0.13-0.59 feet with the L&D 3 pool having a higher MSE of 1.77 feet. These values are found to be generally acceptable for sediment modeling purposes.

## **Sediment and Dredging Model Calibration**

Sediment transport in HEC-RAS can be modeled using a variety of different transport functions, fall velocity equations, bed change options, as well as numerous other calibration parameters. For this modeling effort, multiple different transport functions were investigated initially (Yang, Ackers-White, etc.) but ultimately, the Laursen-Copeland transport function equation was selected for use in the model. The Laursen method (Laursen, 1958) is a total

sediment load predictor developed through experiments and qualitative analysis for grain sizes between 0.011 and 29 mm. Copeland (Copeland, 1989) contributed to the development of the equation to extend the applicability to gravel-sized sediments. The Laursen-Copeland equation showed promising initial results and is the recommended transport equation to use in HEC-RAS for modeling fine grained sediments, outperforming other transport functions in the very fine sand and very coarse silt range (USACE Hydraulic Reference Manual, 2016). The bed sorting method was set to 'Active Layer' of roughly 1 meter in thickness. Rather than specify 3 or 5 different layers in HEC-RAS, a simplistic approach was used where the "active layer" is the portion that is actively transporting and depositing material and the "inactive layer" is the layer below, where sediments are mixed into from the active layer. The fall velocity method was set to the equation developed by Dietrich (Dietrich, 1982) as that method has shown strong results in past studies and was recommended by Dr. Gary Parker as a superior method compared to the other options in HEC-RAS. The bed change option was set to the 'Reservoir Option', which deposits more sediment in the deeper part of the cross-section. This method was more realistic for the series of reservoirs present in the lock and dam system, as opposed to the other options of depositing and eroding sediment uniformly within the movable bed limits or allowing deposition across the entire wetted area uniformly.

Dredging in HEC-RAS is modeled by specifying a station, elevation, width, and time & date of a dredging event at each cross-section in the model. The dredging events were set for July 15 of each year in the model, to represent the entire season's worth of dredging typically occurring over mid-to-late summer. The dredged volume is removed from the system to reflect the standard practice in the Upper Mississippi River of storing dredged material at dredge disposal sites rather than redistributing the material back into the river at a different location.

Modeling the dredging in HEC-RAS based on specified rules is imperfect compared to the subjective decisions that are made in the actual dredging of the system. Channel maintenance is required to maintain the nine foot navigation channel below the Levee Control Profile (LCP). The nine foot channel currently requires dredging to 10.5 feet below the LCP to ensure sufficient draft for barge traffic. However, in actual practice, when dredging does occur the invert is brought to 12 feet below the LCP in order to gain efficiencies in the dredging program (i.e. over-dredge by 1.5 feet so that other locations may be prioritized the following year). In addition to the planned over-dredging, subjective decisions will be made to minimize mobilization of the dredging equipment and to utilize sediment storage sites efficiently. For these reasons, the modeled dredging may not always accurately reflect what actually occurred in the system. However, the model should, on average, do a good job of capturing the total sediment removed through channel maintenance.

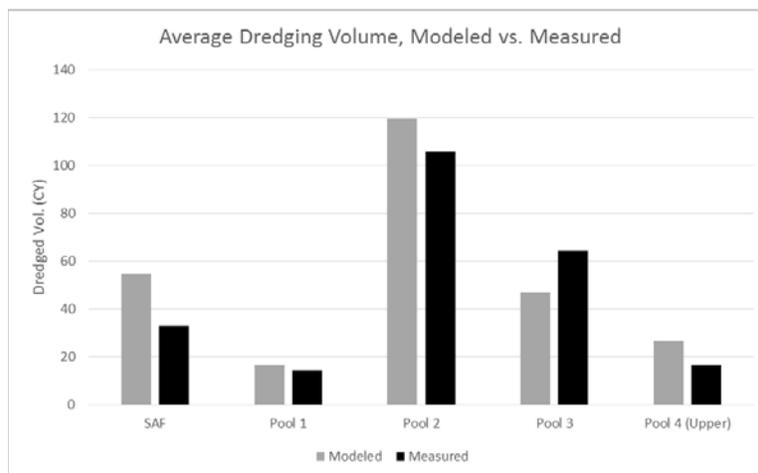
With the modeling methods and dredging events specified, the main calibration parameter used in the model was adjusting the flow-load relationships of the Mississippi River and Minnesota River. Sediment transport, in the model and in reality, reflects the total load of the system which consists of a suspended sediment portion and a bed-load sediment portion. Because the starting flow-load ratings in this model are based on the suspended sediment concentrations, they lack the bed-load sediment estimate, under-predict the total load, and will be adjusted upward during the calibration process. According to the Channel Maintenance and Management Plan, the Upper Mississippi River and tributaries have bed-loads that are between 0 and 40% of the total load, with 10% being the typical value. To account for the bed-load and to achieve calibration, the loads were incrementally increased in the flow-load rating curves until the modeled dredging quantities matched the measured historic dredging quantities. If modeled dredging quantities were low in the St. Anthony Falls Pool and Pool 1, the Mississippi River flow-load curve was increased. If dredging quantities were low in Pools 2-4, the flow-load curve

for the Minnesota River (as the largest contributor of sediment) was increased. For the final calibration the ultimate flow load curves were adjusted to the final curves.

## Results

### Existing Channel Maintenance Practices

Sediment modeling is traditionally very difficult to replicate with high precision and accuracy. Often times, results that are within a factor of two of the measured data are found to be sufficient due to the wide range of variability in sediment data and the complex processes that make up sediment transport. The total dredge quantity modeled in the period from 2008 through 2015 from the Upper St. Anthony Falls Pool through Lake Pepin is 11% higher than the measured volume. Annual quantities for each pool show error sometimes as great as a factor of two, but overall the average modeled dredging quantities compare very well with the average measured quantities. A summary of the average annual dredging volume by pool is shown in Figure 3.



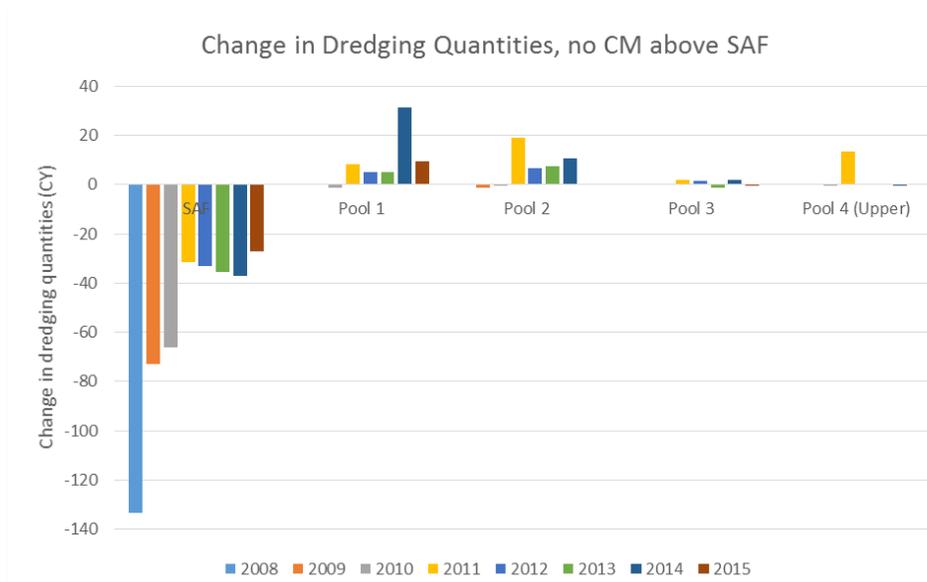
**Figure 3.** Comparison of modeled and measured average annual dredging quantities by pool

With a calibrated sediment dredging model established, the model can be run with various alternatives to show the relative impact the alternative would have to dredging quantities. This calibrated model will be referred to as the base condition model, or the current dredging model. The following sections describe the results of different alternatives to the current channel maintenance plan. These various alternatives will be compared to the base condition model rather than the measured data so that a direct comparison of relative impacts can be made and the residuals between measured and model data will not influence the results.

### Alternative 1 – Eliminate dredging above Upper St. Anthony Falls

The first alternative (Alternative 1) is to eliminate channel maintenance activities above St. Anthony Falls. With the closure of the USAF to navigation that occurred in 2015 as a result of WRRDA 2014, there may no longer be a need to dredge the nine foot channel to boat traffic in the USAF Pool. This alternative is modeled with all dredging activities removed above USAF Lock & Dam. Dredging activities in pools below USAF are modeled using the current dredging plan from the base condition model. The changes in total dredging quantities from the base

condition model of current dredging practices to the Alternative 1 model are shown (summarized by pool and by year) in Figure 4.



**Figure 4.** Comparison of change in annual modeled dredging quantities by pool for Alternative 1

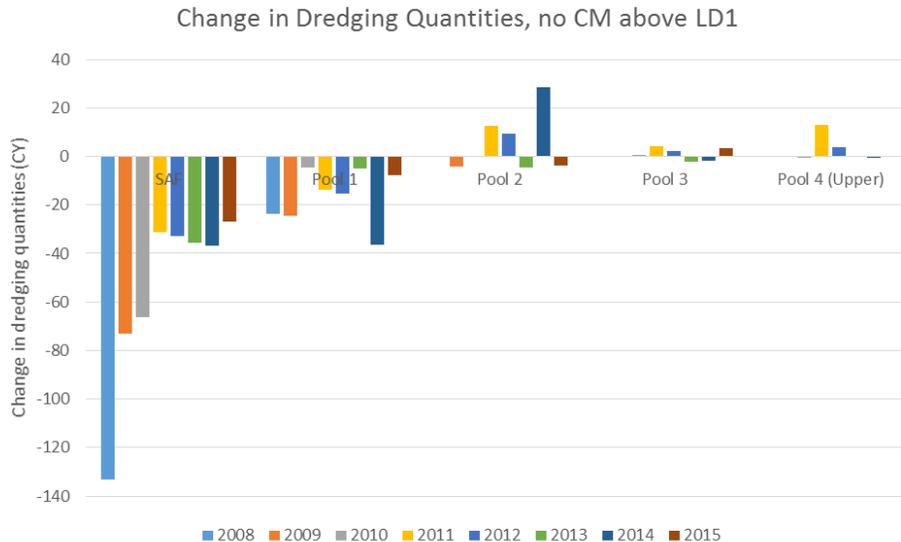
The results of Alternative 1 show increases in dredging for each of the downstream pools, with the greatest increases found in Pool 1. Pools 1 and 2 show positive trends in dredging increases, as well, indicating that the sediment may still be working its way downstream over the 8 year period. Overall, however, the downstream increases in dredging are far less than the total reduction in dredging found in the USAF pool. The relative change in dredging is high for Pool 1 with a 123% increase in total dredging for the pool in Year 8 since the change is implemented for Alternative 1. There is also a strong positive trend in Pool 1, indicating that dredging increases in that pool may continue to be high. The relative change in other pools, however is fairly minimal. The average annual increase to Pools 2, 3, & 4 are 4%, 1%, and 6%, respectively.

## Alternative 2 – Eliminate dredging above Lock & Dam No. 1

The second alternative (Alternative 2) is to eliminate channel maintenance activities above Lock & Dam No. 1, including the elimination of dredging above Upper & Lower St. Anthony Falls. With the closure of the USAF to navigation in 2015, commercial boat traffic in Pool 1 has been minimal in recent years. Data from the Corps of Engineers Lock Performance Monitoring System (USACE LPMS, 2017), shows the typical total lockages for Lock & Dam No. 1 have been reduced from around 1,500 per year to around 1,000 per year, with commercial lockages decreasing from 600 per year to 100 per year.

To represent a scenario where commercial navigation is closed through LD1, Alternative 2 is modeled with all dredging activities removed in USAF Pool and Pool 1. Dredging activities in pools below Lock & Dam No. 1 are modeled using the current dredging plan from the base condition model. The changes in total dredging quantities from the base condition model of current dredging practices to the Alternative 2 model are shown (summarized by pool and by year) in Figure 5.

The results of Alternative 2 show increases in dredging for each of the downstream pools, with the greatest increases found in Pool 2. Pool 2 shows a positive trend in dredging increases, as well, indicating that the sediment may still be working its way downstream over the 8 year period. Overall, however, the downstream increases in dredging are far less than the total reduction in dredging found in the upper pools.



**Figure 5.** Comparison of change in annual modeled dredging quantities by pool for Alternative 2

The relative change in dredging is minimal for Pool 2 with a maximum increase of 14% in Year 7 since the change is implemented for Alternative 2. There is also an increasing trend in Pool 2, indicating that dredging increases in that pool may continue to be high. The relative change in other pools is also minimal. The average annual increase to Pools 2, 3, & 4 are 4%, 2%, and 8%, respectively.

## Conclusions

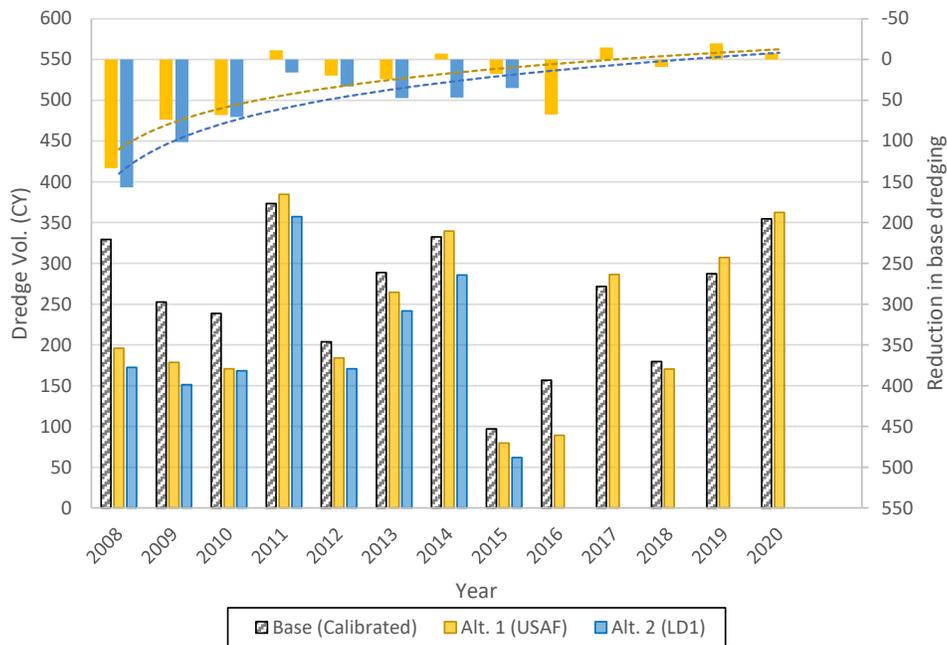
### Comparison of Existing and Proposed Alternatives

Both alternatives, the removal of dredging above Upper St. Anthony Falls and the removal of dredging above Lock & Dam No. 1, result in a net reduction in average dredging volumes over the eight year modeling period. While Alternative 1 results in increased average dredging quantities in Pools 2, 3 & 4 of 4%, 1% and 6%, respectively, the total average dredging for the system shows a net decrease of 15%.

Similarly, Alternative 2 shows increased average dredging in Pools 2, 3 & 4 of 4%, 2%, and 8%, respectively, but a net decrease in average dredging for the system of 24% due to the removal of channel maintenance in Pool 1 and above St. Anthony Falls.

When the total modeled dredging is compared over time (Figure 6), trends in the sediment transport through the system can be identified. In the first 3 years in the model following the implementation of each alternative (2008-2010), the system shows overall reductions in dredging of 29-40% for Alternative 1 and 30-48% for Alternative 2. However, in Year 4 (2011), Alternative 1 shows an increase in total dredging of 3% and Alternative 2 shows only a 4% reduction in dredging quantities compared to the current dredging plan. Toward the end

of the 8 year model period, Alternative 1 again shows a year where dredging quantities exceed the current dredging plan quantities (2014) and both alternatives show less of a reduction in total dredging than in Year 1.



**Figure 6.** Trend in total change in dredging quantities after channel maintenance change. (The columns extending from the bottom of the plot represent total dredged volume (left y-axis) and the columns at the top of the plot represent reduction in base dredging (right y-axis). Note that the Alternative 2 model run was unstable for the future scenario beyond 2015.)

A trend-line for Alternative 1 shows that the expected change in dredging quantities for the system is close to zero by the end of a 10 year period. This might suggest that the system has reached an equilibrium by the end of 10 years and that the total dredging quantities may be net neutral with the current dredging practices. The downstream pools on average, will have slightly higher required dredging to compensate for the lack of a sediment sink above USAF. Alternative 2 shows a similar trend, although it may take longer than a decade to reach equilibrium. Beyond 10 years, the system may expect to be net neutral with the current dredging practices and additional dredging may be required in the downstream pools, on average.

The sediment transport modeling results indicate that eliminating dredging in Pool 1 and/or the USAF Pool will result in significant net reductions in average dredging between the USAF pool and Lake Pepin in the near term. Dredging in Pools 2, 3, and Upper 4 will increase a small amount, however, the reduction in dredging upstream of Pool 2 more than compensates for the downstream increases. Some of the sand not dredged in the USAF Pool and Pool 1 ends up settling out in off-channel areas. However, in the long term, modeling results indicate that once the new equilibrium is reached with each of the alternatives, it is likely that nearly 100% of the new forgone dredging material will end up in the immediate downstream pool. That is, for Alternative 1 most of the dredging increases after 10 years will occur in Pool 1 and for Alternative 2, most increases will occur in Pool 2.

Changes in downstream sediment transport and dredging won't occur immediately, but rather will take a number of years. The model results indicate that the timescale for these changes to

occur may be a decade or two. Aleatory variability in the future hydrology for the system and epistemic uncertainty in the sediment quantities and characteristics lead to high uncertainty in the estimated timeframe for equilibrium, but the model confirms expected trends in sediment deposition with the introduction of each of these alternatives.

### **Model as a tool to investigate sediment trends**

In addition to using the model to assess different channel maintenance management strategies, the model can also be used as a tool to investigate sediment trends in the Mississippi River through Lake Pepin. Numerous studies in recent decades have looked into water quality (Lung & Larson, 1995), rates of deposition (McHenry et al 1980), and sources of sediment (Engstrom et al 2009) in Lake Pepin. This model could be used as a tool to support each of those areas of concern as well as similar fields throughout the Upper Mississippi River. The modeled longitudinal pattern of sediment deposition and particle size change match the measured sediment properties from other researchers (McHenry 1980, Cumulative Effects Report 2000, Engstrom 2009) and they match main channel borings obtained in this reach by the Corps in 2010. This portion of the river, between RM 785 and RM 780, also defines the delta at the upstream end of Lake Pepin. By having this modeling capability to not only capture the sediment budget but to be able to model and predict the grain sizes and locations of sediments, this tool can help with future studies to forecast future water quality and lake capacity concerns for this part of the river.

### **Model as a tool to investigate operational changes**

Recent interest has been sparked to consider even more drastic changes to the navigation system than channel maintenance strategies. The Corps of Engineers has expressed interest in investigating the federal interest in continued operation of the upper three lock & dam structures through a Disposition Study (USACE, 2016). This sediment transport model could be considered, along with numerous other types of models and tools, as one source of information for identifying positive and negative impacts from a change in the operating pools or full removals of dams. The model can coarsely capture the progression of erosion of sediment behind the dam in the case of a removal, but more importantly help to quantify broader impacts to the Mississippi River system through Lake Pepin.

Again, this model would only be one line of evidence in trying to predict the success of such a large scale dam removal project in a highly visible area. With the appropriate amount of additional work and funding, however, this model could prove to be a valuable asset in helping to support or screen-out options to restore the Mississippi River Gorge.

## **References**

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